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¹ Drying-Induced Evaporation of Secondary Organic Aerosol during ² Summer

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6 **S** [Supporting Information](#page-6-0)

⁷ ABSTRACT: This study characterized the effect of drying on ⁸ the concentration of atmospheric secondary organic aerosol ⁹ (SOA). Simultaneous measurements of water-soluble organic 10 carbon in the gas $(WSOC_g)$ and particle $(WSOC_p)$ phases ¹¹ were carried out in Baltimore, MD during the summertime. To 12 investigate the effect of drying on SOA, the $WSOC_p$
13 measurement was alternated through an ambient channel measurement was alternated through an ambient channel 14 (WSOC_p) and a "dried" channel (WSOC_{p,dry}) maintained at ¹⁵ ∼35% relative humidity (RH). The average mass ratio between 16 WSOC_{p,dry} and WSOC_p was 0.85, showing that significant 17 evaporation of the organic aerosol occurred due to drying. The ¹⁸ average amount of evaporated water-soluble organic matter 19 (WSOM = WSOC \times 1.95) was 0.6 μ g m⁻³; however, the 20 maximum evaporated WSOM concentration exceeded 5 μ g

21 m⁻³, demonstrating the importance of this phenomenon. The systematic difference between ambient and dry channels indicates

22 a significant and persistent source of aqueous SOA formed through reversible uptake processes. The wide-ranging implications of

23 the work are discussed, and include: new insight into atmospheric SOA formation; impacts on particle measurement techniques; 24 a newly identified bias in PM_{2.5} measurements using the EPA's Federal Reference and Equivalent Methods (FRM and FEM);

²⁵ atmospheric model evaluations; and the challenge in relating ground-based measurements to remote sensing of aerosol

²⁶ properties.

I. INTRODUCTION

 Organic aerosol (OA) is a ubiquitous and highly abundant 28 component of fine particulate matter.^{[1](#page-6-0)} A majority of this OA is secondary (SOA), even in urban areas where primary emissions 30 are high.^{[2](#page-7-0)} State-of-the-art models are challenged to predict OA 31 concentrations in most locations.^{[3,4](#page-7-0)} Further, model treatment of SOA has derived from laboratory experiments, but there are systematic differences in SOA composition commonly observed 34 between laboratory and ambient studies.^{[5](#page-7-0)} This indicates that significant progress is still required to develop a complete understanding of the sources and formation of SOA. The uptake of water-soluble organic gases into atmospheric liquid water has been identified as a major route for SOA formation that could help to close the model-measurement discrepancy in 40 many locations. 6 ,

 The formation of SOA through the uptake of water-soluble organic gases can occur in cloud and fog droplets and in aerosol liquid water. A key point is that this aqueous SOA formation (aqSOA) involves the uptake of gases that are too volatile to otherwise partition to the aerosol phase without liquid water. There is mounting evidence for the important contribution of aqSOA to the global OA budget. For example, isoprene is the nonmethane hydrocarbon emitted into the atmosphere in the 49 greatest abundance, 8 and liquid water plays a critical role in the

amount of SOA formed from isoprene oxidation.^{[9](#page-7-0)} Several 50 isoprene oxidation products, notably glyoxal and isoprene- 51 derived expoxides (IEPOX), are important aqSOA precur- ⁵² sors.^{[10](#page-7-0)} Neither IEPOX^{[11](#page-7-0)} nor glyoxal^{[12](#page-7-0)} partition to dry particles, 53 but both readily form aqSOA in aerosol water (and presumably ⁵⁴ cloud and fogwater, as well). Based on the high atmospheric 55 emissions of isoprene, IEPOX and glyoxal may be among the ⁵⁶ most abundant SOA precursors on regional and global ⁵⁷ scales. $10,13$

While aqSOA is thought to be an important contributor to ⁵⁹ global OA levels, many uncertainties remain in understanding 60 aqSOA on a fundamental level.^{[7](#page-7-0)} Among the largest 61 uncertainties is the relative contribution of reversible and 62 irreversible uptake processes to aqSOA formation. The initial ⁶³ uptake of water-soluble organic gases into atmospheric water ⁶⁴ occurs through reversible, equilibrium partitioning. We use the 65 terms "reversible aqSOA" and "irreversible aqSOA" to indicate 66 the fate of the dissolved organic gases under conditions of 67 liquid water evaporation. A reversible process implies that ⁶⁸

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 organic compounds taken up into liquid water repartition back to the gas phase upon liquid water evaporation, such as after a cloud cycle or with changes in ambient RH. An irreversible process implies that the same organics remain in the condensed phase with the evaporation of liquid water. The abundance of aqSOA can differ by orders of magnitude depending on which 75 process—reversible or irreversible—is primarily responsible for 76 the uptake of organic gases.^{[14](#page-7-0)}

⁷⁷ Laboratory studies find evidence for both reversible and ⁷⁸ irreversible uptake of glyoxal and IEPOX. Seed particle ⁷⁹ composition is a critical factor that affects the relative split ⁸⁰ between reversible and irreversible uptake. Glyoxal uptake is ⁸¹ enhanced by inorganic aerosol components due to a "salting in" 82 effect.^{[15](#page-7-0)} While some glyoxal remains in the condensed phase 83 under pure glyoxal-water droplet evaporation experiments,^{[16](#page-7-0)} 84 ammonium (NH_4^+) greatly enhances the amount of glyoxal 85 remaining in the aerosol phase under drying conditions.^{[17](#page-7-0)} ⁸⁶ Sulfate appears to play a critical role in catalyzing the 87 ammonium-glyoxal reactions in particles undergoing drying.^{[18](#page-7-0)} ⁸⁸ The relative contributions of reversible and irreversible ⁸⁹ uptake of IEPOX are also strongly affected by inorganic aerosol 90 composition.^{[19](#page-7-0)} Nguyen et al.^{[11](#page-7-0)} observed entirely reversible 91 uptake of $β$ -IEPOX to aqueous sodium chloride (NaCl) or 92 sodium sulfate $\rm (Na_2SO_4)$ particles, while substitution of $\rm NH_4^+$ ⁹³ for Na+ resulted in significant irreversible uptake. There was 94 also an effect from nucleophile substitution, as irreversible β -95 IEPOX uptake to aqueous $(NH_4)_2SO_4$ was an order of 96 magnitude greater than to aqueous NH_4Cl^{11} NH_4Cl^{11} NH_4Cl^{11} These experi-⁹⁷ ments are qualitatively consistent with ambient observations in ⁹⁸ the southeastern U.S., where isoprene-SOA is strongly 99 associated with sulfate. $20,21$ $20,21$ $20,21$ Note that the organic aerosol ¹⁰⁰ composition can also affect the uptake of glyoxal, for example, ¹⁰¹ by altering equilibria and condensed-phase reaction ki-102 netics.^{[22](#page-7-0),[23](#page-7-0)} However, the effect of organic composition on ¹⁰³ IEPOX SOA has not been investigated.

 To date, there are no ambient measurements to quantify or constrain the relative contributions of reversible and irreversible aqSOA. This is important, as it directly affects the impact of 107 aqueous multiphase processes on the total OA budget.^{[16](#page-7-0)} It also has consequences for the fate and lifetime of many organic compounds in the atmosphere. This may have implications for global climate, as well, beyond the general contribution to the OA budget: reactions under conditions of water evaporation can rapidly form aqSOA that contains light-absorbing "brown 113 carbon" species.^{[18](#page-7-0)}

 The purpose of this study was to characterize the behavior of ambient SOA under conditions of drying. A recently developed method was deployed to cycle the measurement of water- soluble organic carbon in the particle phase (WSOC_p) between 118 a dry channel and an unperturbed ambient channel.^{[24](#page-7-0)} A 119 surrogate for SOA, $WSOC_p$ concentrations were unaffected by drying during a previous deployment of the system during the summer−fall transition in the eastern United States.[24](#page-7-0) There was evidence for aqSOA formation during this time, suggesting that the uptake of water-soluble organic gases to aerosol water [24](#page-7-0) occurred through an irreversible process.²⁴ Measurements for the present study were carried out across July and August, a time period that was characterized by higher OA concentrations and warmer temperatures than our previous study, and potential differences in precursor VOC emissions and inorganic 129 aerosol levels. If differences in the behavior of $WSOC_{p}$ were observed due to drying, it would represent the first direct evidence for reversible aqSOA in the atmosphere.

II. MATERIALS AND METHODS

Ambient measurements were carried out in Baltimore, MD ¹³² from 6 July to 15 August 2015. A schematic of the experimental ¹³³ setup is provided in the [Supporting Information \(Figure S1\).](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) ¹³⁴ Details of the experimental setup are described elsewhere. 24 135 Briefly, water-soluble organic carbon was measured in the gas- ¹³⁶ and particle phases $(WSOC_g$ and $WSOC_p$, respectively). 137 $WSOC_{\sigma}$ was measured using a mist chamber (MC) coupled 138 with a total organic carbon (TOC) analyzer (model 900 Turbo, ¹³⁹ GE Analytical).^{[25](#page-7-0)} WSOC_p concentrations were measured using 140 a particle into-liquid sampler (PILS, Brechtel Manufacturing) ¹⁴¹ coupled with the same TOC analyzer used above, according to ¹⁴² the method of Sullivan et al.^{[26](#page-8-0)} All components were housed in a 143 temperature-controlled environmental enclosure (EKTO, Inc.) ¹⁴⁴ placed on the rooftop of the Engineering Building at the ¹⁴⁵ University of Maryland, Baltimore County (UMBC). ¹⁴⁶

To characterize the effect of drying on $WSOC_{p}$, the particle 147 measurement was alternated between a channel maintained at ¹⁴⁸ ambient relative humidity (RH) and a "dried" channel ¹⁴⁹ maintained at ∼35% RH ± 2.6% ([Supporting Information](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) ¹⁵⁰ [Table S1](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf)) using a 3-way valve (Brechtel Manufacturing). The ¹⁵¹ goal for the WSOC_{p,dry} measurement was not to remove all 152 particle bound water, but rather to approximate the lowest RH ¹⁵³ that particles may be exposed to in ambient air during the study ¹⁵⁴ period to simulate "natural" drying processes [\(Supporting](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) ¹⁵⁵ [Information Figure S2](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf)). The dried channel included a silica gel ¹⁵⁶ diffusion dryer, which was made in-house similar to commercial 157 models (e.g., TSI model 3062). WSOC_p losses through the 3- 158 way valve and through the dried channel were evaluated prior ¹⁵⁹ to the start of the sampling period and were found to be ¹⁶⁰ negligible ([Supporting Information Figure S3](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf)). The dryer was ¹⁶¹ replaced daily and its efficiency was checked with an orange ¹⁶² silica gel color-indicator as well as an RH sensor (Omega, RH- ¹⁶³ USB) that measured the RH of air exiting the dryer. The total ¹⁶⁴ residence time at the reduced RH, which includes time inside ¹⁶⁵ the dryer and in the downstream carbon denuder, was ∼7 s. ¹⁶⁶

The fully automated system ran in cycles that were 167 completed every 14 min. The three samples $(WSOC_{g}$, 168 $WSOC_p$ and $WSOC_{p, dry}$) were measured every cycle with 169 sampling times of 4, 5, and 5 min, respectively. These sampling 170 times were averaged to provide a single measurement for each ¹⁷¹ 14 min cycle. Dynamic blanks for both gases and particles were ¹⁷² measured every 2 days when the DI water reservoirs were filled. ¹⁷³ The limits of detection (LOD) were calculated as $3\times$ the 174 standard deviation of the dynamic blanks and were 0.34 μ g-C 175 m⁻³ and 0.30 μ g-C m⁻³ for WSOC_g and WSOC_p, respectively. 176 The relative uncertainty for the WSOC_g measurement is 7%^{[25](#page-7-0)} 177 while that for the WSOC_p measurement is 8% + 0.3 μ g-C 178 $\text{m}^{-3.26}$ $\text{m}^{-3.26}$ $\text{m}^{-3.26}$ The evaporated WSOC_p concentration (i.e., reversible 179 . aqSOA) was calculated as the difference between the $WSOC_p$ 180 and $WSOC_{p,dry}$ concentrations for each cycle. The uncertainty 181 in the reversible aqSOA concentration is the greater of 11% or ¹⁸² 8% + 0.1 μ g-C m⁻³ [\(Supporting Information\)](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf). This is based 183 upon the combined uncertainties of (1) the difference in the ¹⁸⁴ raw TOC analyzer signal between successive dry and ambient ¹⁸⁵ measurements, (2) the PILS air and liquid flow rates, and(3) ¹⁸⁶ the PILS dilution factor. In this case, the blank measurement ¹⁸⁷ does not contribute to the uncertainty in the evaporated ¹⁸⁸ $WSOC_p$ concentration since the same background value is 189 subtracted from the raw TOC data to calculate the ¹⁹⁰ concentrations of $WSOC_{p}$ and $WSOC_{p, dry}$ in air. The 191 uncertainty and the LOD are largely determined by our ability ¹⁹²

Figure 1. (a) Scatter plots of WSOM_{dry} versus WSOM for (a) daytime and (b) nighttime periods. The solid black lines in (a) and (b) represent the linear fits to the data using least-squares regression analysis; the fit parameters are given in each panel.

Figure 2. Histogram of evaporated WSOM. Figure on the lower right corner is a zoom-in of the histogram between 2−5.1 µg m^{−3}. Figure on the upper right corner is a boxplot of the evaporated WSOM concentrations showing median value (horizontal line), 25th and 75th percentiles (lower and upper box values), as well as 5th and 95th percentiles (vertical lines). The dotted orange lines at evaporated WSOM equal to $\pm 0.2 \mu g m^{-3}$ represent the range of the method detection limits.

 to measure a small difference in the TOC concentration of the aqueous PILS sample. The smallest concentration difference that can be distinguished between successive PILS samples is 0.71 ppb-C, which corresponds to an LOD for the evaporated 197 WSOC_p concentration of 0.086 μ g-C m⁻³. This indicates that 198 the LOD for the evaporated WSOC_p concentration is smaller 199 than the LOD for either the $WSOC_{p}$ or $WSOC_{p, dry}$ concentrations. Again, this is due to our method of interpolating between the periodic blank measurements (approximately daily) to estimate the background concen-203 tration for each $WSOC_{p}$ and $WSOC_{p, dry}$ sample, while the evaporated WSOC_p concentration is based upon a difference in the TOC analyzer signal, which is measured for each 14 min 206 cycle. The $WSOC_{p}$ and $WSOC_{p, dry}$ method LODs could be improved by performing more frequent blank measurements, but this was unnecessary for the summertime period: greater ²⁰⁸ than 99% of the WSOC_p measurements were above the 0.3 μ g- 209 C m⁻³ LOD. The WSOC_p concentrations were converted from 210 organic carbon (μ g-C m⁻³) to organic mass (μ g m⁻³) using a 211 WSOM:WSOC ratio of 1.95. This ratio was obtained for a prior ²¹² study carried out during the summertime in the northeastern ²¹³ U.S.^{[27](#page-8-0)} The total organic matter concentration corresponding to 214 $WSOC_{p}$ measurement is denoted as "WSOM" while that 215 corresponding to $WSOC_{p, dry}$ is denoted as "WSO M_{dry} ". 216

III. RESULTS AND DISCUSSION

Effect of Particle Drying on SOA. An overview of the 217 WSOC measurements and the meteorological data for the ²¹⁸ entire study period are summarized in [Supporting Information](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) ²¹⁹

 [Table S3](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf). The WSOC data were segregated into daytime (08:00 to 19:00, local time) and nighttime (20:00 to 07:00, local time) periods in order to characterize different processes 223 that may affect reversible aqSOA formation. [Figures 1](#page-3-0)a and [1b](#page-3-0) show a comparison between the WSOM concentrations measured through the two different channels. During both 226 daytime and nighttime periods, $WSOM_{drv}$ concentrations were systematically lower than WSOM concentrations, indicating that the evaporation of aerosol water also led to the evaporation of condensed-phase organics. The estimated range for the slopes shown in [Figure 1](#page-3-0) were 0.873−0.890 (daytime) and 0.818−0.841 (nighttime) at the 99% Confidence Interval. Such narrow intervals at a high confidence level indicate that these values are tightly constrained, and are statistically different from a value of 1. This is further supported by a comparison of the 235 WSOM_{dry}/WSOM mean ratios, which were statistically differ-236 ent from a $WSOM_{\text{div}}/WSOM$ ratio of 1 during the day and the night [\(Supporting Information Table S2\)](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf).

 238 Similar measurements carried out during a different season^{[24](#page-7-0)} as well as control experiments ([Supporting Information Figure](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) [S3](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf)) showed no difference between WSOM concentrations through the dry and ambient channels. It is also important to note that the channel for the WSO M_{dry} measurement does not employ sample heating to achieve drying, thereby avoiding any organic aerosol evaporation due to a change in vapor pressure. These factors clearly demonstrate that the differences observed in [Figure 1](#page-3-0) were caused by the loss of WSOM due to aerosol water evaporation. Numerous studies have shown that the WSOC_p measurement (converted to WSOM in this study) is a surrogate for SOA in diverse environments.[25](#page-7-0)[,28](#page-8-0)−[31](#page-8-0) This holds when biomass burning influence is low, as is frequently the case in the eastern U.S. during the summer.^{[32](#page-8-0)} Thus, it was highly likely that SOA compounds accounted for the evaporated WSOM observed during the present study.

 The concentration of the evaporated SOA was calculated as the difference between the ambient and dried particle concentrations (WSOM - WSO M_{dry}). [Figure 2](#page-3-0) shows a histogram of the evaporated WSOM. The average amount of 258 evaporated WSOM was 0.6 μ g m⁻³; however, the maximum 259 evaporated value exceeded 5 μ g m⁻³, demonstrating the importance of this phenomenon. Negative evaporated 261 WSOM values indicate measurements where $WSOM_{\text{div}}$ > WSOM. These cases accounted for less than 5% of the total observations, and were below the minimum concentration difference that could be detected for the majority of cases (indicated by the orange dotted lines and the "LOD" label in [Figure 2](#page-3-0)). The systematic difference between WSOM and 267 WSOM_{dry} was observed consistently during day and night periods indicating the important role of aerosol water in summertime SOA formation.

 WSOC Partitioning to Aerosol Water. The paired gas and particle WSOC measurements have provided detailed insight into the formation of aqSOA in aerosol water.^{[6,24,25](#page-7-0)[,33](#page-8-0)} During the present study, the fraction of WSOC in the particle 274 phase $(F_p = WSOC_p/(WSOC_p + WSOC_g))$ showed a strong increase with increasing RH during the night [\(Supporting](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) [Information Figure S5](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf)). The increase in F_p at higher RH levels during the night was statistically significant at the 95% confidence level.

 This RH dependence has been observed in other locations, and is due to the uptake of water-soluble organic gases into 281 aerosol water.^{[25](#page-7-0)} In prior measurements by our group carried out in the fall, we observed irreversible aqSOA formation: the

 $WSOC_o$ taken up into aerosol water remained in the condensed 283 phase with aerosol water evaporation.^{[24](#page-7-0)} For the summertime 284 measurements of the present study, the loss of WSOM back to ²⁸⁵ the gas phase upon drying indicates that some fraction of the ²⁸⁶ aqSOA was formed reversibly. These observations represent, to ²⁸⁷ our knowledge, the first direct atmospheric evidence for ²⁸⁸ reversible aqSOA formation. In a previous study in Mexico ²⁸⁹ City, the observation of reversible aqSOA was hypothesized, ²⁹⁰ based upon strong daytime correlations between $WSOC_p$ and 291 aerosol nitrate. 34 In that study, thermodynamic equilibrium 292 modeling showed that aerosol nitrate rapidly volatilized during ²⁹³ a daytime transition period that was characterized by the ²⁹⁴ evaporation of a significant amount of aerosol liquid water, as ²⁹⁵ well.^{[34](#page-8-0)} However, the volatilization of WSOC_p (i.e., SOA) due 296 to increasing ambient temperature or dilution could not be ²⁹⁷ ruled out, since the temperature increase and boundary layer ²⁹⁸ expansion were coincident with the evaporation of aerosol ²⁹⁹ water. Here, we provide unambiguous evidence for the ³⁰⁰ evaporation of SOA that comes about as a result of particle ³⁰¹ drying. 302

The difference between the current results and our prior ³⁰³ study^{[24](#page-7-0)} is somewhat puzzling. Conditions that promote 304 reversible aqSOA formation were clearly present during July ³⁰⁵ and August (2015), but not during the month of September ³⁰⁶ (2014). Further studies are needed to characterize the actual ³⁰⁷ reason for the differences, which may include precursor VOC ³⁰⁸ emissions, 35 inorganic aerosol composition, $11,17$ and organic 309 aerosol matrix effects. 22 22 22 310

Sources of Reversible aqSOA. The amount of evaporated ³¹¹ WSOM showed a statistically significant increase with ³¹² increasing RH [\(Figure 3a](#page-5-0)). This is highly consistent with the 313 f3 F_p -RH relationship [\(Supporting Information Figure S5a\)](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) and 314 further demonstrates the observation of reversible aqSOA. ³¹⁵ Aerosol liquid water content increases with increasing RH, ³¹⁶ enabling greater uptake of $WSOC_{g}^{23,25}$ $WSOC_{g}^{23,25}$ $WSOC_{g}^{23,25}$ $WSOC_{g}^{23,25}$ $WSOC_{g}^{23,25}$ Some fraction of the 317 $\mathrm{WSOC}_{\mathrm{g}}$ uptake to aerosol water during the summer was 318 reversible, explaining the increase in evaporated WSOM with ³¹⁹ RH. This point is also illustrated in the diurnal profile of ³²⁰ evaporated WSOM during the study period ([Figure 4\)](#page-5-0), which 321 f4 shows that higher concentrations of evaporated WSOM were ³²² observed during the night when RH was generally higher ³²³ [\(Supporting Information Figure S2\)](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf). ³²⁴

There are two additional points that should be noted from ³²⁵ [Figures 3](#page-5-0)a and [4](#page-5-0). First, the evaporation of WSOM was ³²⁶ observed even during daytime periods when the enhanced ³²⁷ partitioning of $WSOC_{g}$ to aerosol water was not pronounced 328 [\(Supporting Information Figure S5b](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf)). This suggests that the ³²⁹ uptake of $WSOC_{\sigma}$ to aerosol water occurred throughout the 330 day, with some uptake proceeding reversibly, but the magnitude ³³¹ of the uptake was reduced during the day given lower aerosol ³³² water content. Second, it is important to note that WSOM ³³³ evaporation was observed even at RH levels between 30 and ³³⁴ 40% ([Figure 3a](#page-5-0)). This suggests that the aerosol in Baltimore ³³⁵ contained liquid water during the summer even at the lowest ³³⁶ RH levels, consistent with observations in other locations.^{[36](#page-8-0)} 337 Khlystov et al. 37 showed that aerosol in Pittsburgh, which is 338 very similar to Baltimore in climatology and aerosol ³³⁹ composition, likely contains water during the summer at ³⁴⁰ these low RH levels, as well. Overall, this observation ³⁴¹ demonstrates that even small amounts of aerosol water (<1− ³⁴² 2μ g m⁻³) can be important for aqSOA. As discussed above, the 343 evaporated WSOM represents a proxy measurement of aqSOA ³⁴⁴ formed through a reversible pathway, so [Figure 3](#page-5-0)a shows that ³⁴⁵

Figure 3. (a) Scatter and box plots of the evaporated WSOM as a function of RH. Data were binned according to the ambient RH: bins were defined as 0−40%, 40−50%, 50−60%, 60−70%, 70−80%, and 80−100%. For each bin, mean (red marker), median (horizontal black line), 25th and 75th percentiles (lower and upper box values), as well as 5th and 95th percentiles (vertical lines) are shown. Numbers at the bottom represent the number of points within each bin. The red dotted line at evaporated WSOM = 0 is for visual reference. The green asterisk on top of bins points to those bins that show a statistically significant increase in comparison to the first bin (i.e., 30−40% RH) at the 95% confidence level. (b) Scatter and box plots of the particulate evaporated WSOM as a function of WSOM concentrations. Data were binned according to the WSOM concentrations: bins were defined as 0−2, 2−3, 3−4, 4−5, 5−6, 6−7, 7−8, 8−9, 9−10, and 10−15 μ g m⁻³. . For each bin, mean (red marker), median values (horizontal line), 25th and 75th percentiles (lower and upper box values), as well as 5th and 95th percentiles (vertical lines) are shown. Numbers at the top represent the number of points within each bin. The red line at evaporating WSOM = 0 is for visual reference.

 aerosol water makes a significant contribution to summertime SOA formation in the study region. In addition to a relationship with RH, the amount of evaporated WSOM increased as the total WSOM concentration increased (Figure 3b). This shows that the source or sources of reversible aqSOA were highly related to those that had a dominant impact on overall SOA formation. In Maryland, as in much of the eastern U.S., it is expected that SOA derives predominantly from biogenic emissions.[32](#page-8-0) Isoprene is the dominant biogenic VOC in this region, and contributes substantially to SOA formation during summer. We hypothesize that the reversible uptake of IEPOX and glyoxal were responsible for the majority of evaporated

Figure 4. Boxplot of the diurnal profile of evaporated WSOM concentrations. For each bin, median values (horizontal line), 25th and 75th percentiles (lower and upper box values), as well as 5th and 95th percentiles (vertical lines) are shown.

WSOM observed during our study. Together, IEPOX and ³⁵⁸ glyoxal are predicted to account for ∼85−90% of aqSOA ³⁵⁹ formed from isoprene oxidation in the eastern U.S. 10 10 10 IEPOX- 360 SOA has been measured in high concentrations in the eastern ³⁶¹ U.S. during summer, $20,21$ and both reversible and irreversible 362 processes contribute to IEPOX uptake in fine particle ³⁶³ water.^{[11](#page-7-0),[19](#page-7-0)} Although direct evidence for glyoxal SOA is elusive 364 due to the lack of specific molecular tracers, models predict ³⁶⁵ significant glyoxal production from isoprene.^{[10](#page-7-0),[38](#page-8-0)} Since glyoxal 366 readily undergoes reversible and irreversible uptake to aqueous ³⁶⁷ particles, $17,23,39$ $17,23,39$ $17,23,39$ it is also a likely contributor to reversible 368 aqSOA concentrations during our study. Further atmospheric ³⁶⁹ measurements, which include speciated VOCs and aerosol ³⁷⁰ chemical components, are needed to test these hypotheses. 371

Implications. These results, which represent the first ³⁷² atmospheric observations of reversible aqSOA formation, ³⁷³ have a number of important implications. It should be noted ³⁷⁴ that these measurements represent a lower-bound on the ³⁷⁵ reversible aqSOA concentrations observed during summer. ³⁷⁶ First, aerosol water was not completely evaporated since the ³⁷⁷ sample was only dried to ∼35% RH, and almost certainly ³⁷⁸ contained residual water. Second, the experimental setup ³⁷⁹ employed a 7 s residence time within and downstream of the ³⁸⁰ dryer before particle collection in the PILS. This amount of ³⁸¹ time far exceeds the equilibration time scale for water ³⁸² evaporation; 40 40 40 however, it may not be enough time for 383 evaporating organic compounds to reach equilibrium.^{[35](#page-8-0),[41](#page-8-0)} If 384 we have systematically underestimated the reversible aqSOA ³⁸⁵ concentration, then our conclusions, and the implications ³⁸⁶ discussed below, may be further amplified.

Atmospheric Chemistry. Laboratory studies find evidence ³⁸⁸ for both reversible and irreversible aqSOA formation. The ³⁸⁹ present results, combined with our prior observations, 24 390 underscore the importance of accounting for both reversible ³⁹¹ and irreversible aqueous-phase processes.^{[7](#page-7-0)} Although reversible 392 aqSOA partitions between the gas and particle phases, like ³⁹³ other semivolatile compounds, it contributes to aerosol effects ³⁹⁴ on health and climate when in the condensed phase. Our ³⁹⁵ results suggest that there are seasonal differences which ³⁹⁶ underlie the split between reversible and irreversible aqSOA ³⁹⁷ pathways, but significant work is needed to characterize these ³⁹⁸ processes on a fundamental level so that they can be effectively ³⁹⁹ implemented into models. Nevertheless, models that treat ⁴⁰⁰ aqSOA formation as entirely irreversible may overstate the ⁴⁰¹

 contribution of this process to the SOA burden in many environments. Conversely, certain instruments may under- estimate the concentration of aqSOA if sample drying occurs as part of the measurement (see detailed [Discussion](#page-3-0) below). This implies, for example, that IEPOX (and glyoxal) may make an even greater contribution to ambient OA concentrations than 408 prior studies have reported.^{[20,21](#page-7-0),[42](#page-8-0)}

 Atmospheric Measurements and Modeling. The present results also have implications for a wide range of atmospheric particle measurements that employ drying. For example, f(RH) measurements compare aerosol light scattering at ambient or humid conditions to light scattering under dry conditions 414 (typically \leq 20% RH).^{[43](#page-8-0)} The difference in light scattering is attributed to aerosol liquid water content; however, our results indicate that evaporated organic aerosol likely contributes to this difference in many environments, as well. Particles also undergo drying in the inlet of two widely used aerosol mass 419 spectrometers^{[44](#page-8-0)}—the Aerodyne aerosol mass spectrometer 420 (AMS)^{45} 420 (AMS)^{45} 420 (AMS)^{45} and various laser-based single particle mass 421 spectrometers^{[46](#page-8-0)}—and are thus subject to the loss of reversible aqSOA during sampling. Indeed, Kampf et al.^{[15](#page-7-0)} inferred the loss of glyoxal taken up reversibly to aqueous particles during sampling with an AMS. The extent of water evaporation in the 425 aerodynamic lens inlet varies from ~50−80%,^{[36,44](#page-8-0)} suggesting that the loss of reversible aqSOA is of the same magnitude. However, systematic studies are needed to characterize this effect under both laboratory and ambient sampling conditions with varied inorganic and organic aerosol composition and liquid water content. Other instruments and methods that implement sample predrying, such as the Hygroscopicity Tandem Differential Mobility Analyzer (HTDMA), are also 433 subject to this effect.³⁶ It is unknown how the present observation may affect OA measurements using thermal-optical analyzers. Factors such as filter handling and storage for offline (laboratory) analysis, or minor variations in filter collection temperature during semicontinuous sampling may impart significant differences in the evaporation of reversible aqSOA by these methods. Note that complete drying (i.e., efflour- escence) is not required for reversible aqSOA to undergo evaporation.

 If certain measurements are subject to WSOM losses during sampling, then that could also affect the evaluation of models, especially those that implement aqueous SOA formation schemes. Models are typically judged in comparison to measurements, which are taken as the "truth", but a systematic bias due to the evaporation of WSOM could lead to incorrect interpretations of model performance. Similarly, the present results may help to reconcile some of the challenges in relating aerosol measurements from the ground with those from remote sensing platforms. There is a known RH dependence in the relationship between aerosol optical depth (AOD) and in situ 453 PM_{2.5} mass,^{[47](#page-8-0)-[49](#page-8-0)} which is thought to be a product of aerosol water. However, the evaporation of WSOM by in situ measurements could also contribute to this observation, since reversible aqSOA contributes to AOD when it is in the particle phase. Our results suggest this effect is important in the eastern U.S., where previous discrepancies between surface and remote 459 sensing measurements have been identified.^{[48,50](#page-8-0)}

460 Regulation and Health. These results also have implications ⁴⁶¹ for the EPA's Federal Reference and Federal Equivalent 462 Methods (FRM and FEM) for measuring $PM_{2.5}$. The FRM 463 and FEM employ drying to $35 \pm 5\%$ RH to eliminate the 464 contribution of aerosol water to measured PM mass.^{[51](#page-8-0)} These

methods are likely subject to a bias from the evaporation of ⁴⁶⁵ WSOM that accompanies the filter or sample conditioning. ⁴⁶⁶ Based on a comparison to measurements at the closest ⁴⁶⁷ Maryland Department of the Environment ambient monitoring ⁴⁶⁸ station to UMBC (Oldtown, ∼ 10 km distance), we infer an ⁴⁶⁹ average bias in hourly $PM_{2.5}$ concentrations measured with 470 FEM method 127 (BAM monitor) of −5%. However, the ⁴⁷¹ largest inferred bias during the study period, calculated as the ⁴⁷² ratio of the evaporated WSOM concentration to the measured ⁴⁷³ PM_{2.5} concentration, exceeded -20% . The negative bias 474 indicates that the FEM (and FRM) systematically under- ⁴⁷⁵ measure the actual concentration of $PM_{2.5}$ due to the 476 evaporation of WSOM that occurs with the evaporation of ⁴⁷⁷ particle-bound water. Our results suggest that this phenomenon ⁴⁷⁸ is important in the eastern U.S. during summer, but significant ⁴⁷⁹ work is needed to fully understand the spatial scale and ⁴⁸⁰ magnitude of this effect. The potential for a systematic bias in ⁴⁸¹ FRM and FEM determinations of PM_2 , has implications for 482. compliance with the National Ambient Air Quality Standards ⁴⁸³ (NAAQS). While the evaporation of other semivolatile ⁴⁸⁴ compounds, notably ammonium nitrate, has been well ⁴⁸⁵ documented for the FRM and FEM, 52 the present results 486 identify a new bias not previously considered. The evaporation ⁴⁸⁷ of reversible aqSOA during sampling may also have ⁴⁸⁸ implications for health effects, beyond the general contribution ⁴⁸⁹ to PM mass. Future studies should assess the potential of this ⁴⁹⁰ reversible aqSOA to generate reactive oxygen species (ROS), ⁴⁹¹ which is thought to underlie many of the adverse health ⁴⁹² outcomes that result from exposure to $PM.⁵³$ $PM.⁵³$ $PM.⁵³$ This recom- 493 mendation is based upon the strong correlation between ⁴⁹⁴ reversible aqSOA and the overall $WSOC_p$ in this study [\(Figure](#page-5-0) 495 [3](#page-5-0)b), and close associations between WSOC_p and ROS observed 496 in diverse locations,^{[53](#page-8-0)} including in the southeastern U.S.^{[54](#page-8-0)} 497
■ ASSOCIATED CONTENT 498

\bullet Supporting Information 499

The Supporting Information is available free of charge on the 500 [ACS Publications website](http://pubs.acs.org) at DOI: [10.1021/acs.est.5b06002](http://pubs.acs.org/doi/abs/10.1021/acs.est.5b06002). 501

Figures S1−S5, Tables S1−S3, and additional method ⁵⁰² details ([PDF\)](http://pubs.acs.org/doi/suppl/10.1021/acs.est.5b06002/suppl_file/es5b06002_si_001.pdf) 503

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